

## ORIGINAL ARTICLE



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# Thermal tolerance in Pacific salmon: A systematic review of species, populations, life stages and methodologies

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## Abstract

Aquatic systems are warming and exceeding upper thermal limits (UTLs) for many fish species, yet understanding how they inform resilience to climate change is challenging. Using Pacific salmon (*Oncorhynchus* spp.) as a model, we conducted a systematic review involving 168 papers investigating UTL in five species. We found considerable variation in UTL among species, within species and across life stages; largely attributed to methodological approaches (e.g.  $CT_{max}$ /ULT, Aerobic/Cardiac Scope, Thermal Migration Barriers, Rearing Mortality, Thermal Preference/Avoidance). Given that each method has strengths and weaknesses owing to logistics, time scale and ecological realism, we offer a new framework for assessing vulnerability to warming that stresses the importance of considering UTL metrics in the context of intended use (i.e. the development of management guidelines, projections of future persistence and survival) and what aspect (physiological or behavioural) of thermal response a metric investigates. Comparing studies with identical UTL approaches revealed that within species, UTL was higher for populations historically encountering higher temperatures—suggesting local thermal adaptation. Within populations, UTL differed across the lifecycle, being highest in fry and lowest in embryo and migrating adults, but this was not universal. For spawning Pacific salmon, UTL has not been assessed and few studies have examined estuarine and marine stages. Ultimately, this data gap limits the life-history thermal perspectives that can be drawn and may indicate a broader gap for all fishes, given that Pacific salmon are among the most well-studied species. Our framework illustrates the inherent and methodological inconsistencies in UTL and offers a guide for how thermal limits can best be used to assess the warming tolerance and vulnerability of fishes.

## KEYWORDS

behaviour, climate change, methodology, mortality, *Oncorhynchus*, upper thermal limits

## 1 | INTRODUCTION

Vulnerability of fish to high temperatures is primarily a product of three main factors: thermal tolerance (range of temperatures in which a species can persist), thermal exposure (temperatures

experienced by a species throughout its lifecycle) and adaptive capacity (the ability for a species to adapt to changes in environmental conditions through evolution and phenotypic plasticity) (Williams et al., 2008). Yet a scheme for assessing thermal tolerance that incorporates these elements is lacking. In this paper, we describe a

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new framework for assessing thermal tolerance that focuses on the three primary elements of a species' vulnerability to warming and outlines how current thermal limits can best be used for assessing climate warming vulnerability in fish.

Few groups of species have been studied thoroughly enough in a thermal tolerance context to develop such a framework, with the exception of the Pacific salmon (*Oncorhynchus* spp., Salmonidae). As a model system, this group of species is ideal for this task as they possess a narrow optimal temperature range (Farrell, 2009), they have a limited ability to extend upper thermal tolerance (Muñoz et al., 2014, 2018) and are exposed to a wide range of climate change impacts due to their anadromous migratory behaviours (Flemming & Jensen, 2002). Natal homing and localized selective pressures have facilitated the development of finely tuned local adaptations that allow for higher fitness in native environments and have given rise to thousands of genetically distinct populations of Pacific salmon (Adkison, 1995; Fraser et al., 2011; Taylor, 1991). Specifically, intra-specific variation in thermal tolerance has been found to correlate with historical temperature experiences (Farrell et al., 2008) and suggests that these fish possess some adaptive potential to warming (Zillig et al., 2021). However, many populations are already imperilled throughout their historical range, and it remains unknown whether Pacific salmon can adapt their thermal tolerance to match the pace at which temperatures have and will continue to increase (Crozier et al., 2019; Zhang et al., 2019).

The reported upper thermal limits (UTLs) for Pacific salmon can vary quite extensively. The extent of the variation in reported UTL and the amount that can be attributed to natural variability in thermal tolerance versus variability introduced by the methodological details remains unknown (Baird et al., 2018; Clark et al., 2011; Poletto et al., 2017). As a result, assessing vulnerability and predicting how Pacific salmon may respond to continued warming will be highly dependent on the temperature threshold selected for any given analysis. Further, the variation in UTL that may emerge from research using different methods, life stages or populations will make it challenging for decision-makers to apply this information to management or conservation actions. Despite the presence of several reviews on the governing effects of temperature on the growth, production and survival of Pacific salmon (McCullough, 1999; McCullough et al., 2001; Myrick & Chech, 2004; Richter & Kolmes, 2005; Zillig et al., 2021), a more comprehensive framework is needed to outline how thermal metrics can be better used in predicting the response of Pacific salmonids to climate change, and our intention is that this framework can be broadly adopted by researchers examining thermal tolerance in other species.

The objectives of this review are to (1) categorize the different methodologies employed to define UTL in Pacific salmon, (2) assess how research into the governing effects of temperature on these species has changed over time, (3) classify the extent and sources of variation in UTL, (4) examine how life stages, populations and species vary in UTL and identify where research is limited and (5) discuss how UTL can be used to develop a thermal framework to assist in predicting future persistence and survival of fish in a warming world.

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## 2 | METHODS

### 2.1 | Data collection

To comprehensively collect studies that report UTL metrics for Pacific salmon, we conducted a systematic literature search in two

databases: Web of Science and Aquatic Sciences and Fisheries Abstracts. The search included all papers published from January 1900 to June 2020 and relevant papers published after that point were added to the review manually. We used a keyword-based search using the following query combinations:

1. Focal species: sockeye salmon OR Chinook salmon OR coho salmon OR pink salmon OR chum salmon
2. Temperature synonyms: thermal\* OR temperature OR warming OR climate OR heat OR thermoregulat\*
3. Keywords to encompass the diverse approaches used to assess thermal tolerance in ecological and physiological research: tolerance OR threshold OR limit OR resistance OR resilience OR stress\* OR surviv\* OR lethal OR sublethal OR optim\* OR maxim\* OR behavior OR infection OR disease OR recovery OR distribution OR migrat\* OR exercise

The keywords were formatted to include all relevant iterations of a root word using an asterix (e.g. migrat\* to encompass migratory migration migrating, migrate). We also scanned the reference sections of all relevant studies to check for additional publications (peer-reviewed or grey literature) that had not been captured through our initial keyword-based literature search.

## 2.2 | Data organization and analysis

The scope of this review was limited to the five anadromous and semelparous North American species of Pacific salmon (sockeye salmon [*Oncorhynchus nerka*], Chinook salmon [*Oncorhynchus tshawytscha*], coho salmon [*Oncorhynchus kisutch*], chum salmon [*Oncorhynchus keta*], pink salmon [*Oncorhynchus gorbuscha*]). Only papers that reported novel numerical thermal metrics for these species of Pacific salmon were included in this review. We excluded papers that looked at the relationship between temperature and growth unless a scope curve for growth was developed and a temperature at which growth could no longer occur was identified. Studies that assessed the cumulative effects of temperature and chemical exposure were excluded and research on the interaction between temperature and disease were only kept if they looked at how temperature reduced the resilience of the fish to the infection or pathogen. We read through the abstract, methods, results and discussion sections of all relevant papers.

Each study was classified by species, life stage, stock, study type, decade, methodology and corresponding temperature metric. Temperature metrics were reported as the mean value (with standard deviation or standard error of the mean if reported by the authors). Distributional studies were reported as the maximum value recorded during the study. We identified seven major study types (Swimming/Physiology, Incubation and Rearing, Capture/Count, Holding, Capture Simulation, Tagging and Model) that have been used to assess the effects of temperature on salmon function and behaviour (see Table S1).

We further identified seven distinct methodologies (Table 1) that have been used to assess UTL for Pacific salmon stemming from both physiological and ecological fields of research. Our goal in classifying each methodology is to identify sources of variation in thermal limits reported and provide insight on how these values can be used when making projections about the impacts of climate change on Pacific salmon. We also made comparisons among different methodologies based on their relative ecological relevance which we defined as: representative of a realistic scenario for what the test fish experience in nature and reflective of the temperatures at which fitness-promoting activities can no longer be performed. We focused on the following methodologies (described in detail in Section 3.2): Critical thermal maximums ( $CT_{max}$ ), upper incipient lethal temperatures (UILT), preference/avoidance, thermal migration barriers, scope for activity/aerobic scope and cardiac function. We selected these methodologies for further analysis as they are most commonly employed in the literature when quantifying thermal optimums and upper limits of Pacific salmon across species and life stages. Finally, in order to further illustrate how reported UTL differ due to intraspecific variation (i.e. species, population, life stage) and methodology-based variation (both within and among methodologies), we also draw upon four case studies where we compare reported temperature limits from a number of studies collected in this review.

## 3 | RESULTS AND DISCUSSION

### 3.1 | General trends in the literature

There were 168 peer-reviewed studies and technical reports included in this review that spanned across life stages and species of Pacific salmon (see Table S1). Of these, 20 were not found in our literature search but were included as additional relevant sources. The majority of studies were focused on the freshwater juvenile life stage (57%), followed by freshwater adult migrants at various stages of maturity (38%), smolts (3%) and lastly immature ocean fish (1.8%) (see Table S1). The limited attention given to marine portion of the lifecycle is not unexpected given that during the marine phase, salmonids do not experience temperatures near their known UTLs (Morita et al., 2010). Freshwater life stages are of greater concern with respect to the impact of climate-induced warming, especially for adults, as they possess finite endogenous energy with which to complete spawning migrations (Cooke et al., 2008). Further, reproductive life stages (spawning and embryonic) are identified as a critical thermal bottleneck in lifecycle of fishes (Dahlke et al., 2020). Chinook (39%) and sockeye (38%) salmon were the most commonly studied species followed by coho (28%), pink (10%) and chum salmon (7%). Pacific salmon that spend the most time in freshwater (i.e. stream-type Chinook, lake-type sockeye and coho) are believed to be more vulnerable to climate change compared to those that migrate to the ocean more immediately after emergence (i.e. pink, chum and ocean-type Chinook salmon; Grant et al., 2019).

The higher potential vulnerability to climate change of sockeye, coho and stream-type Chinook salmon may partially explain why these species have historically received greater attention in the literature in terms of assessing UTL metrics. Pink and chum salmon exhibit higher rates of straying than the other species which may reduce their vulnerability to freshwater habitat changes because greater flexibility in spawning location may allow for higher rates of reproduction in years where conditions in natal streams are not ideal (Gilk et al., 2004; Keefer & Caudill, 2014; Quinn, 1993). Given that abundances of pink and chum salmon are at all-time highs in the North Pacific Ocean, (Ruggerone & Irvine, 2018), and their commercial fisheries value now surpasses the other species of Pacific salmon (Gislason et al., 2017), it is more important than ever to conduct further research into the UTL of the various life stages of pink and chum salmon. Upper thermal tolerance metrics obtained from the literature, by species and life stage, are included in Tables S2–S5.

### 3.2 | Identification and comparison of methodologies

Physiological and ecological research into thermal tolerance of Pacific salmon has given rise to numerous temperature metrics with varying definitions and endpoints (Bates & Morley, 2020). However, the thermal limits proposed vary extensively depending on the

experimental design and represent a range of responses such as thermal resistance, acclimation and evolutionary adaptation (Bates & Morley, 2020). Therefore, the variation in methodological details and the thermal responses that various approaches reflect must be considered when using such metrics to improve management designs and predict future vulnerability to climate change. The following section reviews the four primary methodological approaches that have been employed to assess upper thermal tolerance and discusses their benefits and potential limitations, especially in regard to their ecological relevance. We felt that comparing the methodologies in this manner would situate them within the broader ecological context of thermal performance and limitations since, in nature, fish can experience temperature variations at multiple temporal scales from diurnal to seasonal fluctuations as well as across the life-cycle. For example, Spring and summer-run Chinook in the Columbia naturally experience diel fluctuation of ~2–5°C (Keefer et al., 2019). Further, during the migratory adult life stage of Koeys river sockeye salmon, average temperatures have been found to fluctuate from 12 to 18°C (Atlas et al., 2021). These natural fluctuations can either mediate or exacerbate threats posed by changes in the hydro climatology of salmon-producing watersheds. Such changes refer to both shifts in climate trajectories (presses) and extreme events (pulses) (Harris et al., 2018). Therefore, long-term changes in temperature and flow must be considered in combination with shifts in the frequency and intensity of extreme events to improve the detection of climate change impacts on biological systems (Harris et al., 2018).

TABLE 1 Behavioural and physiological methodologies employed to quantify upper thermal tolerance in Pacific salmon.

	Methodologies	Description
Behavioural	Preference/avoidance	Temperature-based observations on areas where fish most densely congregate and/or actively avoid
	Thermal migration barrier	Temperatures at which migrating up or downstream becomes challenging (e.g. slower passage rates, cessation of migration, decreased survival)
	Temperature limits to distribution	Presence or absence surveys linked to temperature profiles to determine temperature-based limits to distribution
	Temperature-driven changes in behaviour	Determining temperatures at which obvious changes in behaviour occur (e.g. increased used of off-channel thermal refugia, decreased ability to avoid predation)
Physiological	Scope for activity ( $T_{opt(AS)}$ , $T_{crit(AS)}$ , $T_p$ , $T_{ABT}$ , $T_{arr}$ )	Quantification of optimal and critical temperatures for performing activities beyond basic maintenance
	Scope for growth	Quantification of optimal and critical temperatures for growth through the development of a temperature-dependent scope for growth curve
	Upper incipient lethal temperature (UILT)	Abrupt transfer from acclimation to exposure temperature. Resistance times to an acute temperature increase is assessed
	Critical thermal maximum ( $CT_{max}$ )	Gradual heating from acclimation temperature to the temperature at which test fish lose equilibrium
	Incubation and rearing optimums and limits	Quantification of optimal and critical temperatures for embryonic and early juvenile life stages through incubating at a range of temperatures and observing differences in survival between groups
	Critical temperature ( $T_{crit}$ )	Temperature below which no mortality occurs
	Cellular/tissue response	Molecular approach to identifying critical temperature thresholds in fish (e.g. induction of heat shock proteins, transcriptomics)

### 3.2.1 | $CT_{max}$ /UILT

#### Background

$CT_{max}$  and UILT are identified through acclimating/acclimatizing organisms to a given temperature and involve time and temperature as primary test variables (Becker & Genoway, 1979; Beitinger et al., 2000). However, these approaches were developed independently and possess two critical differences.  $CT_{max}$  involves a continuous warming from the acclimation to exposure temperature with loss of equilibrium as the endpoint (Abe et al., 2019; Baird et al., 2018). UILT involves a rapid transfer from acclimation to exposure temperature with death as the endpoint (Becker & Genoway, 1979; Brett, 1952).

#### Ecological relevance

Traditional  $CT_{max}$  and UILT studies generate consistently high thermal limits compared to other methodologies and possess the least ecological relevance (Figure 1). First, the rapid changes in temperature that occur in  $CT_{max}$  studies, and the abrupt changes used to develop estimates for UILT do not reflect a realistic scenario for the thermal experience of Pacific salmon in nature (Becker & Genoway, 1979). Second, for  $CT_{max}$  in particular, the lethal limit obtained is highly influenced by the acclimation temperature and rate of warming and does not necessarily reflect any genetic or adaptive characteristics of the test subjects (Kingslover & Umbanhowar, 2018; McKenzie et al., 2020; Terblanche et al., 2007). During continuous warming, a lag between body temperature and ambient water temperature can occur if rate of warming is too rapid and there is a risk of partial acclimation to higher temperatures than intended when warming too slowly (Becker & Genoway, 1979; Desforges et al., 2023). Further, the mechanisms that lead to loss of equilibrium remain unknown and may not be ubiquitous across species or warming rates, which makes it challenging to determine the cause of variation in  $CT_{max}$  among species or explain why the resulting values are modified by acclimation temperature (Lefevre et al., 2021). Third, organ systems central to fitness-promoting activities often decline at far lower temperatures than those quantified by the  $CT_{max}$  or UILT approaches. This mismatch between tissue function and loss of righting response or death calls into question both the functional utility and ecological realism of these approaches and limits their use beyond being a simple relative measure of thermal tolerance across populations, species or organisms (Muñoz et al., 2014). However, efforts to develop models that allow for comparison across  $CT_{max}$  and UILT trials with varying warming protocols are being undertaken (Jørgensen et al., 2021; Ørsted et al., 2022). This research offers a promising framework for improving comparative assessments of thermal tolerance across populations where critical thermal limits have been obtained. Even then, comparing across studies is challenging as UTL can be altered by the methodological details of any given experiment (Terblanche et al., 2007). Comparing studies that employ the  $CT_{max}$  methodology, in particular, is difficult if no standardized rate of warming has been established. Additionally,

a recent study on upper thermal tolerance in brown trout (*Salmo trutta*) found that intraspecific variation in  $CT_{max}$  did not have any clear relationship with migration timing or strategy, growth or predation vulnerability (Desforges et al., 2021). Their results provide further evidence that  $CT_{max}$  values are not indicative of how temperature affects a fish's ability to perform fitness-promoting activities. However, this metric is widely used across a broad range of fish species and is repeatable, easy and relatively low cost to perform (Beitinger et al., 2000). Therefore, we do not discourage use of this metric, rather we encourage both careful interpretation and combined use of metrics to improve the ecological relevance of the thermal thresholds used in future management schemes.  $CT_{max}$  values may be the only thermal metric available for a population, and when this is the case, these values should be considered in proper context—are there  $CT_{max}$  and  $T_{crit}$  values for other populations that exhibit similar life histories and exist in similar climates? This information could then be used to infer that ecologically relevant thermal thresholds are 'x' degrees lower than what the  $CT_{max}$  suggests.

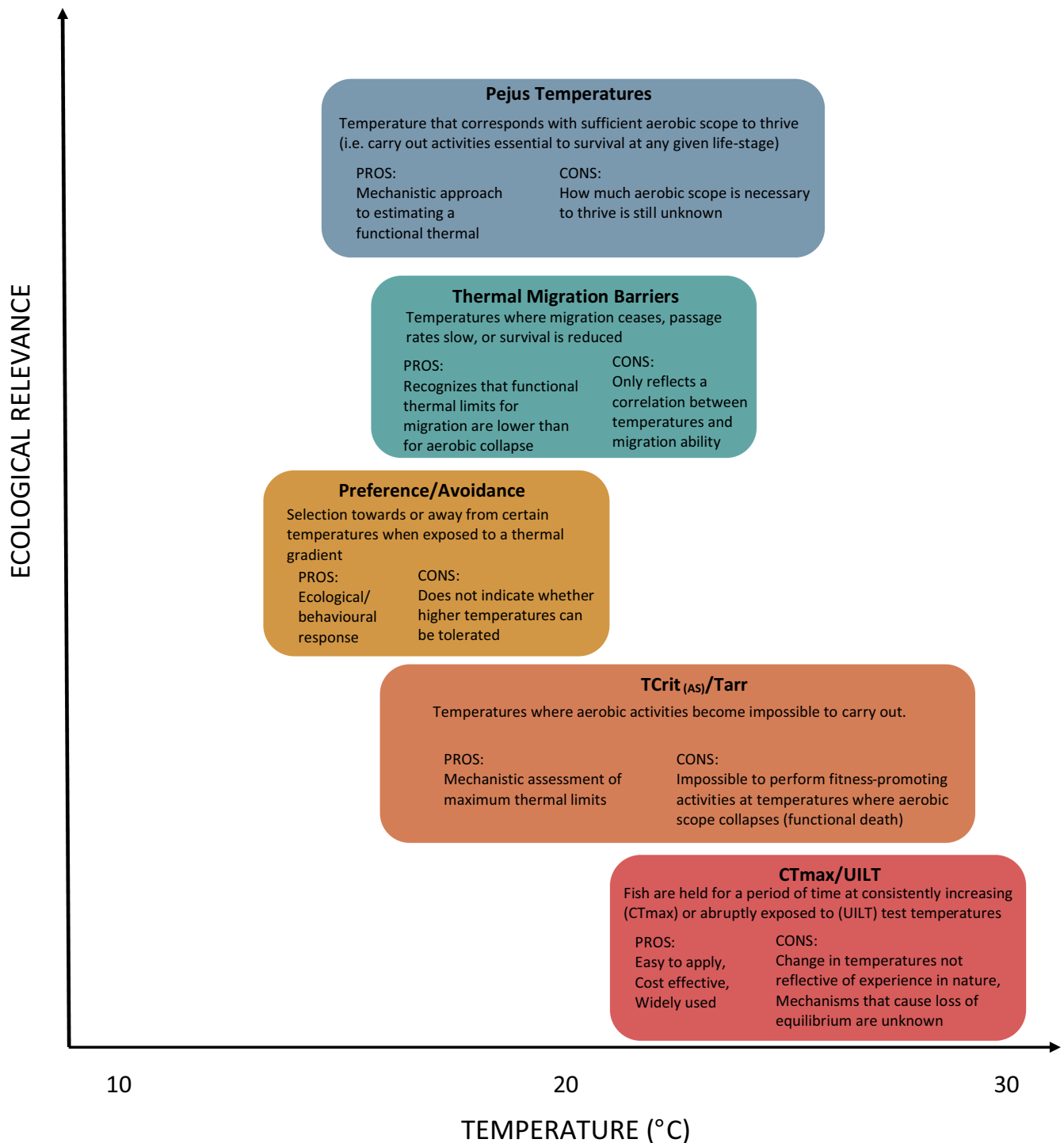
### 3.2.2 | Preference/avoidance

#### Background

Fish have evolved to be thermally selective based on temperatures that optimize physiological, ecological and reproductive performance (Brett, 1971). As a result, identifying temperatures that fish actively avoid or prefer became popular throughout the 1970s and 1980s as understanding this temperature-based selective process was integral to incentivizing fish to move towards or away from certain areas (Coutant, 1987). This was especially beneficial for attaining more desirable yields in fisheries (Coutant, 1987). These studies can be performed in both laboratory- and field-based settings. Field-based experiments often utilize tagging technologies to identify areas where fish congregate more densely as well as areas where fish are unlikely to be found and correlate these movements with water temperatures in those areas (Armstrong et al., 2016). Vertical gill netting and trawl surveys have also been employed to define temperature preferences for Pacific salmonids (Crowder & Magnuson, 1982). Laboratory-based experiments are performed in tanks where an artificial thermal gradient is created and the movement and congregation of fish in the water column is observed over some set period of time (see Christensen et al., 2021 for details).

#### Ecological relevance

The preference/avoidance approach consistently reports some of the lowest UTL among the various methodologies compared in this review (Figure 1). This may be partially due to the fact that temperatures that fish actively avoid only suggests a preference for lower temperatures. While temperature preferences can indicate a behavioural manifestation of an individual's physiological response to its surrounding environment (Christensen et al., 2021; Huey, 1991), they do not indicate whether fish are capable of tolerating higher temperatures.



**FIGURE 1** Conceptual Figure illustrating the relative ecological relevance of the critical thermal maximum (CT<sub>max</sub>), upper incipient lethal temperature (UILT), critical temperature for aerobic scope (T<sub>crit(AS)</sub>), arrhythmic temperature (T<sub>arr</sub>), avoidance, thermal migration barrier and Pejus temperatures (T<sub>p</sub>) methodologies for quantifying thermal limits of Pacific salmon. Widths and placement of the boxes reflect the breadth of temperatures for each category.

Furthermore, the majority of these studies are laboratory-based, which creates further complications for relating the experimental observations to the experience of the test species in the wild. In reality, there are numerous environmental factors aside from temperature (e.g. predators, food, salinity, dissolved oxygen levels) that can incentivize fish to select one area over another (Eliaison & Farrell, 2016).

### 3.2.3 | Thermal migration barriers

#### Background

Studies that aim to identify temperature limits to up and downstream migration are largely field-based and dominated by tagging studies (Hallock et al., 1970; Keefer, Caudill, et al., 2008; Keefer,



Peery, & Heinrich, 2008; MacDonald et al., 2010). Another common experimental design uses fish counts at dams and weirs to link slower passage rates or limited passage with a thermal barrier for migration (Gonia et al., 2006). Multiple limiting effects of temperature on migration have been assessed in the literature, including delayed migration or slowing of migration rates (Caudill et al., 2007; Crossin et al., 2008; Gonia et al., 2006; Hodgson & Quinn, 2002; Hyatt et al., 2003; Keefer, Caudill, et al., 2008), inability to reach natal spawning grounds (Connor et al., 2019; Keefer, Peery, & Heinrich, 2008) and cessation of migration (Hallock et al., 1970; MacDonald et al., 2010).

#### Ecological relevance

The thermal migration barrier methodology reports similar thermal limits to avoidance studies (Figure 1). This approach demonstrates a correlation between temperature and the ability for fish to migrate upstream and only applies in the smolt and adult migrant life stages. However, there are numerous other factors that reduce an individual's capacity to migrate successfully (e.g. accelerated loss of endogenous resources, pathogen virulence) that can operate independently or interactively with temperature (Martins et al., 2011). There is also a large degree of variation in UTLs reported using this approach likely due to different definitions for upper thermal migratory limits: temperatures when (i) successful migration declines (% of fish that no longer reach spawning grounds, this % threshold varies across studies), (ii) migration ceases (little to no fish detected at receivers further upstream or observed at fish fences), (iii) passage rates slow significantly (time between detections at receivers or fish fences along the migration route increases). This is problematic as variation in reported thermal limits may be less a function of inherent variation in thermal tolerance and more a product of how the thermal limit is defined.

### 3.2.4 | Scope for activity/aerobic scope

#### Background

For aquatic ectotherms like Pacific salmon, thermal limits can be determined by insufficient aerobic capacity to support essential activities (Pörtner & Farrell, 2008). This theory developed when Fry (1947) demonstrated the controlling influence of temperature on a fish's 'scope for activity'. Scope for activity (aka aerobic scope) is the difference between maximum metabolic rate (MMR, the peak metabolism possible) and standard metabolic rate (SMR, the minimum metabolism required to sustain life). Aerobic scope is the amount of energy available to fish to perform any activities beyond basic body maintenance (e.g. swim, digest a meal, compete for mates or territory and escape a predator). Since this seminal study was conducted, research into the temperature dependence of aerobic scope has been built upon and used to determine both optimal ( $T_{opt(AS)}$ —the temperature where aerobic scope is greatest) and critical temperatures ( $T_{crit(AS)}$ —temperature where aerobic scope is zero) for many species and populations of fish. However,

the mechanisms that triggers the decline in maximum oxygen consumption and thus, in aerobic scope, above the  $T_{opt(AS)}$  remains a contentious topic. Pörtner (2001) hypothesized that aerobic scope becomes limited with increasing temperature due to limitations of organ systems to deliver oxygen to tissues (termed the oxygen and capacity limited thermal tolerance hypothesis, OCLTT). Data from migrating adult sockeye salmon support this hypothesis (Eliason et al., 2011; but see counter arguments from other fish species; Lefevre et al., 2021). In sockeye salmon, the decrease in aerobic scope at high temperatures (MacNutt et al., 2006) was associated with a collapse in cardiac performance (Eliason et al., 2011; Eliason, Clark, et al., 2013; Eliason, Wilson, et al., 2013). As a result, recent research into the limiting effects of temperature on cardiac performance has given rise to cardiac-based optimal ( $T_{ABT}$ —the temperature above which maximum heart rate becomes limited with increasing temperature) and critical temperatures ( $T_{arr}$ —the temperature at which heart rate becomes arrhythmic). These values tend to yield temperatures that parallel the  $T_{opt(AS)}$  and  $T_{crit(AS)}$  (Casselman et al., 2012; Muñoz et al., 2015).

In addition to optimal and critical temperatures for aerobic and cardiac scope, there have been a few studies over the past two decades that aim to determine the highest temperatures where Pacific salmon can maintain sufficient aerobic scope to successfully carry out spawning migrations (Eliason et al., 2011; Farrell et al., 2008). These studies assume that a functional temperature limit for successful upriver migration lies somewhere between the  $T_{opt(AS)}$  and the  $T_{crit(AS)}$ , however, the precise temperature will vary depending on the difficulty of the spawning migration and on the individual's maximum aerobic capacity (Farrell et al., 2008). Eliason et al. (2011) referred to such a temperature as the Pejus Temperature ( $T_p$ ) value and suggests that the amount of aerobic scope required for adult migrants to complete spawning migrations is 80%–90% of total aerobic capacity. However, data on temperature dependence of aerobic scope in Pacific salmon are limited (absent for many species and populations), making it very difficult to discern how much aerobic scope is actually required for adults to reach spawning grounds.

#### Ecological relevance

The estimates for optimal and critical temperatures for aerobic and cardiac scope offer a more mechanistic understanding of the controlling effects of temperature on an organism's ability to function than other methodologies employed in the literature. However, similarly to  $CT_{max}$  and UILT thresholds (Figure 1), fitness-promoting activities such as predator avoidance and upriver migrations are impossible to perform at the temperatures where these critical limits occur (i.e.  $T_{crit(AS)}$ ). Additionally, most studies on the temperature dependence of cardiac and aerobic scope are conducted in controlled laboratory settings and prove difficult to relate to an individual's experience in nature. There are also several methodological issues that must be considered when comparing across studies (Lefevre et al., 2021). Specifically, the duration of thermal exposure and the range of acclimation

temperatures varies across studies, which can influence the shape of the aerobic scope curve (Lefevre et al., 2021). When assessing cardiac capacity in relation to temperature,  $T_{ABT}$  and  $T_{arr}$ , experiments are typically performed on anaesthetized and pharmacologically stimulated fish (to ensure they reach max heart rate), which may impact the results in ways that are irrespective of temperature. Further heating rates across aerobic and cardiac scope studies also vary which can impact the acclimation response. Therefore, comparisons across studies should be made with caution. Further, an inability to assess excess post-exercise oxygen consumption and its effects on acclimation means that only aerobic processes are being considered when assessing thermal tolerance (Lee, Farrell, Lotto, Hinch, & Healey, 2003). Lastly, when investigating the temperature dependence of aerobic scope, RMR and MMR are measured and calculated at several experimental temperatures. Since AS is not evaluated over a continuous temperature gradient,  $T_{opt(AS)}$  and  $T_{crit(AS)}$  are typically estimated by fitting a polynomial regression onto the data.

For adult migrants,  $T_p$  estimates are more ecologically relevant (Figure 1) because they reflect the fact that upriver migrations require more than a minimum amount of aerobic capacity. If upriver migrants require 80%–90% of their total aerobic scope to migrate upstream (Eliason et al., 2011), that would correspond to lower temperatures than those at which complete aerobic collapse ( $T_{crit(AS)}$ ) or whole animal incapacitation ( $CT_{max}$ ) occurs. However, it is important to recognize that not all species or life stages utilize the full extent of their aerobic capacity (Kunz et al., 2018; Lowe & Davison, 2006). It remains unknown how much absolute aerobic scope is necessary for juvenile life stages to thrive (e.g. to swim, digest, grow, compete for space, forage). Therefore, while this approach may be useful for migrating adult salmonids, it may not be the best tool for assessing how temperature affects fitness in juvenile life stages where less of their total aerobic capacity is required to carry out activities essential to survival. Due to the potential non-uniformity of the use of full aerobic capacity across life stages and among species, it can be difficult to relate reductions in aerobic scope to overall fitness in the wild (Lefevre et al., 2021).

### 3.2.5 | Case studies: Natural variability and variability within- and among-UTL approaches

There is some degree of natural variability in temperature tolerance among species, populations and life stages which can partially be attributed to different historical thermal experiences (Angilletta et al., 2008) and the length and difficulty of spawning migrations (Crossin et al., 2004; Eliason et al., 2011; Farrell et al., 2008). Our analysis also indicates a high degree of variation in thermal limits reported depending on the methodology employed and the differences in experimental details within a given methodology. In the following section, we explore four case studies to demonstrate the inherent and methodological sources of variability in estimates of optimal temperatures for performance and UTL.

#### *Interspecific variation in UTL metrics: Comparing $T_{opt(AS)}$ across different species of anadromous Pacific salmon*

The data extracted for this case study comes from select research that analysed the impacts of temperatures on optimal metabolic performance using the Scope for Activity/Aerobic Scope methodology (Table 2). We pooled data from four studies (Eliason et al., 2011; Eliason, Wilson, et al., 2013; Farrell et al., 2008; Lee, Farrell, Lotto, MacNutt, et al., 2003) to assess how the average  $T_{opt(AS)}$  for multiple populations of sockeye salmon may differ from other species of Pacific salmon. Table 2 illustrates how the  $T_{opt(AS)}$  varies considerably between these four species with an ~12°C difference in optimal temperatures when comparing a coastal winter-run population of coho salmon and a coastal fall-run population of pink salmon. Interestingly, the  $T_{opt(AS)}$  for sockeye, coho and chum salmon coincide with the historical average temperature experiences during upriver migration (Abe et al., 2019; Eliason et al., 2011; Farrell et al., 2008; Lee, Farrell, Lotto, MacNutt, et al., 2003). However, for pink salmon, the same metric seems to be more reflective of historical maximum temperatures (Clark et al., 2011). It is possible that these differences may be caused by variation in life-history strategies yet, evidence is lacking to provide concrete reasoning for the differences in  $T_{opt(AS)}$  among species of Pacific salmonids.

#### *Intraspecific variation in UTL: Comparing $T_{crit(AS)}$ for adult migrants in seven populations of Fraser River sockeye salmon*

Eliason et al. (2011) and Eliason, Wilson, et al. (2013) were selected to demonstrate the possible extent of intraspecific variation among populations of Pacific salmon from the same region as they are the only studies published to date that develop optimal and critical temperatures for the greatest number of populations from a Pacific salmon species at the same life stage, using the same experimental design across populations. These papers developed aerobic scope curves for seven populations of Fraser River sockeye salmon in British Columbia, Canada under the expectation that intraspecific variation in  $T_{crit(AS)}$  is associated with the difficulty of the upriver migration and historical environmental temperatures. Fraser River sockeye salmon populations co-migrate and thus can be divided into broad run timings based on when they re-enter the river (Early Stuart, Summer Run, Fall Run). Early Stuart sockeye enter the Fraser River in early July, encountering high river flows and warming temperatures as they approach distant spawning grounds. Summer Run populations (Chilko, Quesnel, Gates, Nechako) encounter peak temperatures when they enter the Fraser and moderate river flows. Fall Run populations (Weaver and Harrison) encounter cool temperatures, low flows and have the shortest, least difficult migration. Figure 2 illustrates the correlation between a migration difficulty index (migration slope = migration elevation/migration distance) and the  $T_{crit(AS)}$  and  $T_{opt(AS)}$  estimates for the populations assessed in these two studies.  $T_{crit(AS)}$  is more variable (range: 20.8–29.4°C) than  $T_{opt(AS)}$  (range: 14.5–17.2°C) among populations and  $T_{crit(AS)}$  is more strongly correlated with migration slope. As such, these studies suggest that populations with less challenging, cooler migrations may be less thermally tolerant and thus, more vulnerable to climate-induced increases in freshwater temperatures. Additional studies on chum salmon (Abe



et al., 2019), Chinook salmon (Zillig et al., 2021), sockeye salmon fry (Chen et al., 2013) similarly have discovered that salmon populations differ in thermal tolerance. Collectively, these data suggest that Pacific salmon have a remarkable capacity to respond to their environment via genetic adaptation and/or phenotypic plasticity.

#### *Life stage-specific UTL: Comparisons among embryo, fry and adult migrants in sockeye salmon*

In addition to both inter- and intraspecific variation, research also suggests that thermal metrics differ depending on life stage. To demonstrate this third source of inherent variation in thermal tolerance, Figure 3 illustrates the differences in *relative* thermal tolerance for nine populations of sockeye salmon native to the Fraser River Basin and highlights the finely tuned local adaptations in thermal tolerance during freshwater life stages. Embryo data were extracted from a study by Whitney et al. (2013) assessing thermal tolerance of nine populations during embryonic development. Among the populations used in their research, Chilko and Gates sockeye had the most dramatic declines in embryonic survival when incubated at 16°C relative to 14 and 10°C groups. They suggested that the stark reduction in survival may be due to the historically cool spawning and incubation temperatures experienced by these populations, while the Scotch and Adams populations, with the superior thermal tolerance, incubate in the warmest environments. Chen et al. (2013) developed  $CT_{max}$  values for the same nine populations of sockeye salmon in the fry life stage after incubation under three different thermal regimes (10, 14 and 16°C). Populations did differ in  $CT_{max}$ , although this was strongly related to body size. In general, the absolute differences in  $CT_{max}$  across populations were minimal at this life stage, particularly when compared to a common body size.  $T_{opt(AS)}$  and  $T_{crit(AS)}$  values for migrating adult fish were estimated by Eliason et al. (2011) who discovered that the Chilko population (Summer migrant) had exceptionally high upper thermal tolerance and are more resilient to warming, whereas the Weaver population (Fall migrant) had the lowest thermal tolerance and appeared to be the most vulnerable population to warming during this life stage.

Taken together, these studies suggest that populations are strongly locally adapted to their environmental conditions at both the embryo and adult stage. They also demonstrate that there is considerable intraspecific variation in thermal tolerance and that relative thermal tolerance differs across the lifecycle. For example, the Chilko population had the lowest thermal tolerance at the embryo stage but the highest thermal tolerance at the adult stage. As

such, fisheries managers and other decision-makers should avoid extrapolating thermal tolerance metrics across populations or across the life cycle.

We also compared *absolute* thermal tolerance across the lifecycles of three populations of sockeye salmon, to identify potential thermal bottlenecks (Dahlke et al., 2020; Pörtner & Farrell, 2008) using data extracted from Whitney et al. (2013), Chen et al. (2013), Eliason et al. (2011) and Eliason, Wilson, et al. (2013). The current paradigm suggests that thermal windows are narrowest for embryos and spawners and widest during the juvenile life stage (Dahlke et al., 2020; Pörtner & Farrell, 2008). In this case study, absolute thermal tolerance increased as salmon age from embryos to fry, as predicted. However, three different trends were observed between the transitions from fry to adult. Populations displayed a decrease (Weaver), minimal change (Gates) and increase (Chilko) in thermal tolerance from the fry stage to the migrating, sexually maturing adult stage. Chilko sockeye, in particular, provide an interesting case where the thermal tolerance window is widest for migrating adults, contrary to this existing paradigm. This highlights how bottlenecks in thermal tolerance may differ across populations. Therefore, trends in the way thermal tolerance may change across the lifecycle cannot be generalized. While the degree of life stage-specific variation shown in this case study might be somewhat confounded by the differences in experimental details between studies, the data compiled in Figure 4 is the most detailed information available on life stage-specific UTL across populations. All these methodologies (embryo survival,  $CT_{max}$ ,  $T_{crit(AS)}$ ) reflect temperatures where death is imminent and were performed in controlled laboratory environments. While it would be ideal to only compare across studies that employ the same methodologies, studies that look directly at life stage differences in thermal tolerance are largely absent from the literature. Therefore, the examples provided here represent the best life stage comparison available in the literature on thermal tolerance in Pacific salmon. A recent alternative proposed by FitzGerald et al. (2020) employs a modelling approach where life stage-specific thermal tolerance is standardized by assessing the likelihood of survival to the succeeding life stage at a given temperature.

We found extremely limited data on the upper thermal tolerance of alevin, smolt and marine life stages or the fully mature spawning life stage (Figure 4). There was only one published study that identified a  $CT_{max}$  threshold for Chinook salmon in the alevin life stage (Del Rio et al., 2019). Since alevins have limited mobility and therefore are unable to behaviourally thermoregulate (Myrick & Chech, 2004),

**TABLE 2**  $T_{opt(AS)}$  estimates for sockeye, coho, pink and chum salmon (Abe et al., 2019; Clark et al., 2011; Eliason et al., 2011; Eliason, Clark, et al., 2013; Farrell et al., 2008; Kraskura et al., 2020; Lee, Farrell, Lotto, MacNutt, et al., 2003).

Species	Sockeye	Coho	Chum	Pink
$T_{opt(AS)}$ average (°C)	15.89	12.35	15.80	21.00
$T_{opt(AS)}$ range (°C)	14.50–17.20	8.50–18.00	14.00–17.60	–
Mean migration Temperature (°C)	~16.2	~7.1	~11.1	~16.9
Maximum migration temperature (°C)	~22	~12	~19	~21

Note: Ranges are given when optimal temperatures for aerobic scope have been assessed in more than one study: sockeye  $N=9$  (six populations), coho  $N=2$  (one population), chum  $N=2$  (two populations).

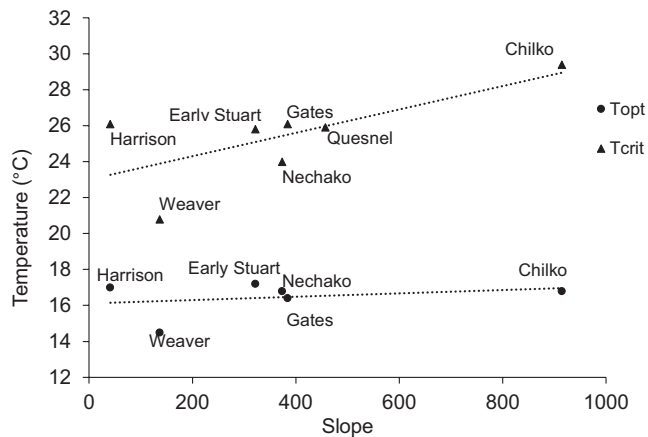
further research into their upper thermal tolerance is necessary. Several studies on smolt and marine life stages have developed temperature preference estimates (Eliason et al., 2017; Hinke et al., 2005) and distributional limits (Azumaya et al., 2007; Welch et al., 1995) but only one study has assessed the UTL in Chinook smolts (Baker et al., 1995). The absence of data on thermal tolerance in mature spawning adults is also a concern given that warming in spawning streams is expected to cause thermal stress in spawning adults (FitzGerald et al., 2020). Further, there is good evidence that for mature spawning adults, thermal tolerance is likely to be lower than has been found for less mature adult migrants. Thermal tolerance is proposed to decrease in reproductively mature adults due

to the increased oxygen demand of gamete production and maintenance (Pörtner & Farrell, 2008). Hruska et al. (2010) assessed physiological changes of individual spawning adults from the Weaver Creek population and found that rapid senescence was associated with increasing concentrations of plasma cortisol and lactate to a point that indicated extremely high levels of anaerobiosis. Relatively high plasma lactate concentrations due to a heightened mismatch in oxygen supply and demand is associated with collapse in thermal tolerance in migrating adult sockeye (Eliason, Wilson, et al., 2013).

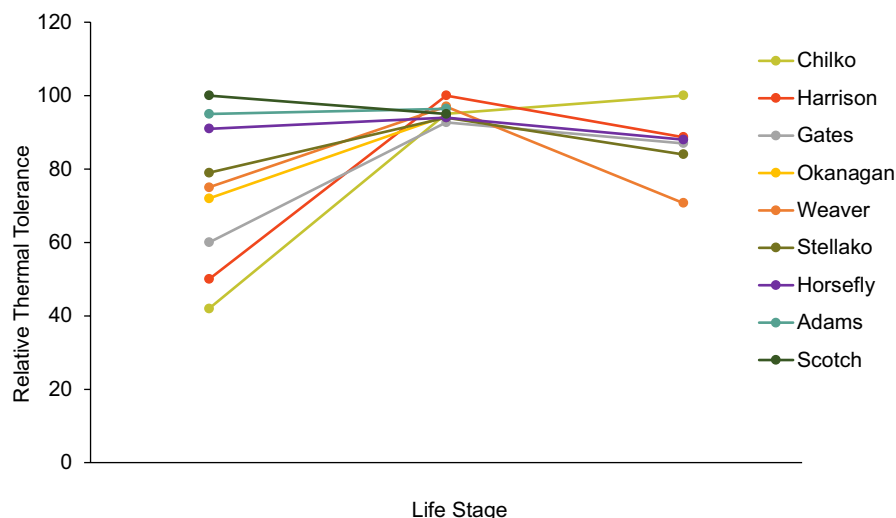
#### Variation in upper thermal tolerance within- and among-methodological approaches

Becker and Genoway (1979) generated  $CT_{max}$  values for juvenile coho salmon under two acclimation temperatures and five different heating rates. In their study, upper thermal tolerance varied by  $\sim 4.5^\circ\text{C}$  when comparing  $CT_{max}$  values for a test group acclimated to  $5^\circ\text{C}$  and heated at a rate of  $1^\circ\text{C h}^{-1}$  compared to a fish acclimated to  $15^\circ\text{C}$  and heated at a rate of  $60^\circ\text{C h}^{-1}$  (Figure 5). Further, the absolute change in  $CT_{max}$  became greater at faster heating rates indicating that in addition to a systematic difference, there may also be a multiplicative effect on the  $CT_{max}$  value derived as heating rate increases. Their results provide an excellent illustration of how traditional  $CT_{max}$  values can be altered significantly by two primary components of the  $CT_{max}$  methodology: the rate at which the water is heated and the initial acclimation temperature. Variation across studies in one or both of these components makes comparison extremely difficult due to the variable impacts that acclimation treatments and heating rates can have on  $CT_{max}$  estimates.

On a broader level, variation in UTLs is also prevalent among different methodologies. The variation in upper thermal tolerance values between  $T_p$ ,  $T_{crit(AS)}$  and upriver migration methodologies indicates that some of the variation in reported upper thermal tolerance limits can also be attributed to the differences between approaches. One



**FIGURE 2** Correlation between migration slope (migration elevation/migration distance) and critical ( $T_{crit}$ ) and optimal ( $T_{opt}$ ) for Aerobic scope for Fraser River adult migrating sockeye salmon populations ( $T_{crit}$  correlation:  $y = 0.0065x + 22.99$ ,  $R^2 = 0.49$ ,  $p = .08$ ;  $T_{opt}$  correlation:  $y = 0.0009x + 16.11$ ,  $R^2 = 0.08$ ,  $p = .81$ ; data from Eliason et al., 2011; Eliason, Wilson, et al., 2013).



**FIGURE 3** Intraspecific variation in relative upper thermal tolerance across the sockeye salmon life cycle. Relative thermal tolerance adapted from critical temperatures for aerobic scope ( $T_{crit(AS)}$ ) in adults (Eliason et al., 2011), critical thermal maximums ( $CT_{max}$ ) for fry (1g) reared at  $10^\circ\text{C}$ , (Chen et al., 2013) and estimates of 60% survival (derived from survival curves for embryos incubated at 10, 14 and  $16^\circ\text{C}$ ) for embryos (Whitney et al., 2013). Relative thermal tolerance for each life stage was generated by relating each population to the population with the highest upper thermal limit (UTL):  $100 - (\text{population max UTL} - \text{population} \times \text{UTL}) / \text{population max UTL}$ .

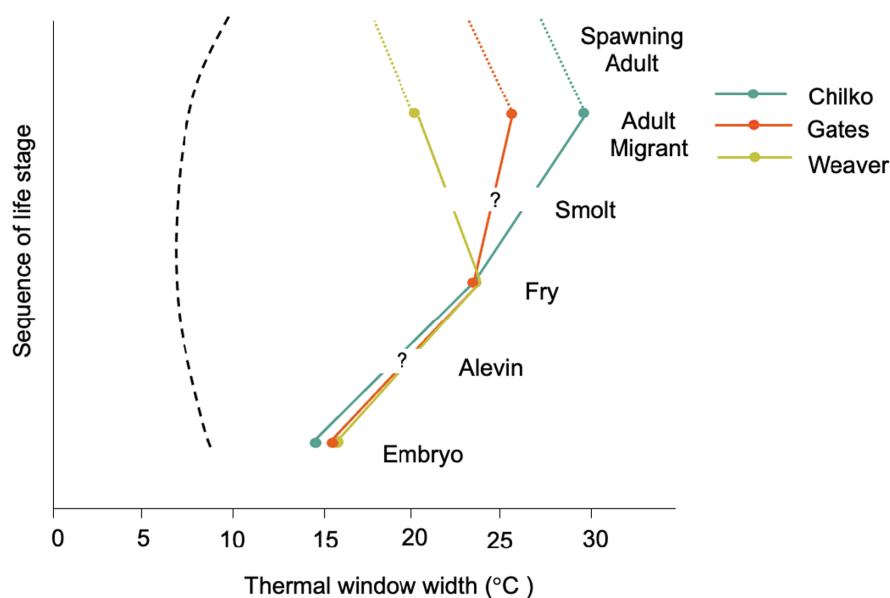
major reason for this variation is that different methodologies define the UTL in different ways. The upper thermal tolerance of adult sockeye salmon from the Early Stuart population has been assessed using two methodological approaches: a 'Thermal Migration Barrier' and 'Scope for Activity/Aerobic Scope' (Table 2). MacDonald et al. (2010) investigated the impacts of environmental conditions (e.g. temperature and discharge) on spawning migration mortality and determined that 18.5°C represented a temperature threshold that explained some of the discrepancy between lower-river and spawning escapement estimates for Early Stuart sockeye salmon. Eliason et al. (2011) employed the Scope for Activity/Aerobic Scope methodology to develop  $T_{crit(AS)}$  and  $T_p$  estimates for the same population. In this case study, the thermal limit to upriver migration and the  $T_p$  estimate correspond closely (18.5 and 19°C respectively). However, the  $T_{crit(AS)}$  was ~7°C higher than the two other more conservative estimates. This reflects the fact that  $T_p$  and Thermal Migration Barrier metrics have both generally set the thermal limit as the temperatures above which successful migration becomes challenging whereas,  $T_{crit(AS)}$  defines the UTL as the temperature where complete aerobic collapse occurs and is not an ecologically relevant metric.

### 3.3 | A framework for assessing vulnerability to warming in the climate change era

An essential component of successful conservation and management in the climate change era is to identify species and populations

that are most vulnerable to climate change and develop actions to mitigate this threat (Bottrill et al., 2008; FitzGerald et al., 2020). In order to comprehensively determine a fish species' vulnerability to warming, there are three main factors to consider: The thermal tolerance (temperature range in which a species can thrive), the thermal exposure (the temperatures that a species experiences through its lifetime) and the adaptive capacity (ability for a population or species to extend their upper thermal tolerance through phenotypic plasticity and/or adaptive evolution) and the pace at which that adaptation occurs (Figure 6).

As researchers begin to recognize the extensive differences in both thermal tolerance and exposure at different life stages in fishes, another major question arises: at which life stages are fish most vulnerable to climate warming? An analysis of thermal bottlenecks in fish species identifies both spawning adults and embryonic development as the most sensitive life stages for many fish species with respect to climate-induced warming of aquatic environments (Dahlke et al., 2020). The authors did not report any thermal ranges for spawning adult Pacific salmon, nor did they account for population-specific differences in thermal tolerance which is a potential limitation to their study. However, their work demonstrates that even if some species exhibit high thermal tolerance at one point in their lifecycle this may not be indicative of overall resilience to future warming. Our investigation into the UTL for Chilko sockeye indicates low thermal tolerance during the juvenile life stage despite superior aerobic and cardiac capacity at high temperatures during the adult migrant life stage. These differences in thermal tolerance



**FIGURE 4** Life stage-specific variation in absolute upper thermal tolerance for Fraser River sockeye salmon populations Weaver, Gates and Chilko based on comparable critical temperatures: critical temperatures for aerobic scope ( $T_{crit(AS)}$ ) in adults (Eliason et al., 2011), critical thermal maximums ( $CT_{max}$ ) for fry (1 g) reared at 10°C (Chen et al., 2013) and estimates of 60% survival in embryos (from survival curves for embryos incubated at 10, 14 and 16°C; Whitney et al., 2013). Dotted lines illustrate hypothetical upper thermal tolerance in mature spawning adults and the black dashed line represents a hypothetical lower temperature tolerance limit. The range of temperatures between the upper and lower tolerance limits represents the thermal window width for a given life stage. The embryo and reproductive adult life stages are predicted to have the narrowest thermal window widths (Dahlke et al., 2020; Pörtner & Farrell, 2008), which is likely true for Weaver, though not for Chilko. Despite considerable research attention on thermal tolerance in sockeye salmon, gaps in knowledge remain.

during a critical period in the lifecycle provides an excellent example of how policies surrounding conservation and management of Pacific salmon must account for the fact that within a single population, there may be specific life stages where greater vulnerability to warming exists. Management at lower levels of biological organization (population and life stage) will promote genetic diversity and allow for greater adaptive and evolutionary potential to future environmental conditions (Zillig et al., 2021).

While there is clear evidence of population and life stage-specific differences in thermal tolerance (Figures 3 and 4), very limited research has been conducted into how sex and size affect UTL. Some studies have demonstrated that smaller individuals, within a life stage, may be better able to withstand high temperature exposures than larger individuals (Clark et al., 2012; Daufresne et al., 2009). Additionally, a recent and comprehensive literature review of sex-specific mortality in adult Pacific salmon revealed that females perish at over twice the rate of males when migration conditions are difficult, and in particular, when rivers are warm (Hinch et al., 2021). The causes of these high levels of female migration mortality are numerous including energy depletion, cardiac collapse, physiological stress and immune issues, and there is evidence that spawning ground sex ratios are being affected by this phenomenon (Hinch et al., 2021). Given these findings, it is imperative that future thermal tolerance research examine the effects of sex and size on all life stages.

Estimating the impacts of climate change based solely on UTL can lead to an underestimation of the negative influences of warming. Throughout their lifecycles, Pacific salmonids will encounter a multitude of biotic and abiotic stressors that may cumulatively affect survival. For example, significant increases in pathogen susceptibility and inability to recover from capture events have been observed when fish are faced with higher temperatures (Gale et al., 2011, 2014; Raby et al., 2015, 2016; Teffer et al., 2018, 2019). Therefore, the interactive effects of these cumulative stressors should be considered alongside reported UTLs when developing thermal guidelines for management and conservation of these species.

While quantifying upper temperature limits is an important component to understanding how cold-water fishes may respond to climate warming, thermal tolerance alone does not determine vulnerability to climate warming. Effective conservation and management of Pacific salmon will also require estimates of current and future thermal exposure at the population and life stage levels. Estimates of water temperatures in rivers across space and time, an understanding of how these species move through river networks and quantifying potential thermal exposure are all required to develop an understanding of current and future vulnerability to warming (FitzGerald et al., 2020). Migratory salmonids complete phases of their lifecycles in different habitats and the thermal tolerance of a specific population in any given life stage can vary extensively (Figures 3 and 4). Moreover, thermal exposure differs depending on which part of the lifecycle an individual is in and can be modulated by shifts in behaviour (e.g. changes in run-timing). Therefore, a thorough understanding of how thermal exposure compares to the UTL

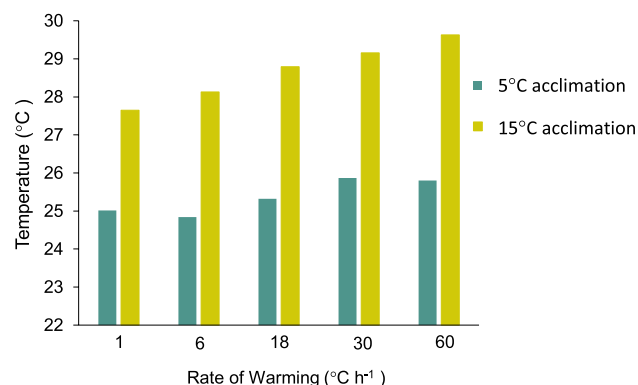
proposed for a population at any point in their lifecycle is another essential component of designing effective management schemes and determining the potential for Pacific salmonids to withstand future warming (Figure 6).

There have been several attempts to relate UTL to current and potential future exposure through the development of concepts such as thermal safety margins (TSM). TSM is an index of the relative proximity of an organism's current thermal experience to its UTL and can be used to compare vulnerability across populations and species (Pinsky et al., 2019). This is usually done by subtracting the known  $CT_{max}$  for a population from the highest current temperatures experienced (Sunday et al., 2014; Vinagre et al., 2019). However, there are limits to the use of TSM in making predictions about how populations may fare under future warming as sublethal temperatures are major drivers of mortality and extirpation before TSM's reach zero (Pinsky et al., 2019). Further,  $CT_{max}$  reflects a maximum tolerance level and does not possess the same ecological relevance as other metrics (e.g.  $T_p$ , thermal migration barriers). Therefore, developing TSM's that utilize a more functional range of UTLs (e.g.  $T_p$ , thermal migration barriers, temperature where growth is limited) may provide a more realistic understanding of a populations' vulnerability to warming. For example, a recent study on the population-specific thermal tolerance of juvenile rainbow trout suggests using the  $T_p$  rather than the  $CT_{max}$  provides a more ecologically relevant estimate of a population's warming tolerance (Anlauf-Dunn et al., 2022).

Beyond developing a knowledge based on the UTL and the thermal exposure of Pacific salmonids, an understanding of the potential for thermal adaptation is another key component in predicting the consequences of climate change for these species. Both developmental and intergenerational plasticity can play a major role in an individuals' survival and fitness as well as population persistence (Figure 6). For example, Burt et al. (2012a, 2012b), found that, parental identity had a significant influence on embryonic survival, particularly under a high-temperature incubation treatment (16°C). Further, fry incubated at higher temperatures have been shown to maintain higher maximum upper thermal tolerance limits in the subsequent life stage (Chen et al., 2013).

In response to warming, fish populations can (a) alter behaviour to avoid unfavourable conditions (e.g. Minke-Martin et al., 2018; Roscoe et al., 2010), (b) acclimate through phenotypic plasticity to withstand high temperatures over short time periods (minutes, hours, days) or (c) adapt through natural selection to increase thermal tolerance at the gene level (Anttila et al., 2014). However, given the current scenarios for future warming in the Pacific Northwest, non-genetic responses such as phenotypic plasticity are unlikely to be sufficient to withstand climate warming for many populations (Crozier, Hendry, et al., 2008; Crozier, Zabel, & Hamlet, 2008). Therefore, evolutionary adaption of thermal tolerance should be considered when making predictions about long-term resilience in native environments (Chen et al., 2017). As temperatures continue to rise, it is probable that selection will act in favour of thermally tolerant phenotypes as long as there is heritable variation in these traits. However, whether thermal adaptation

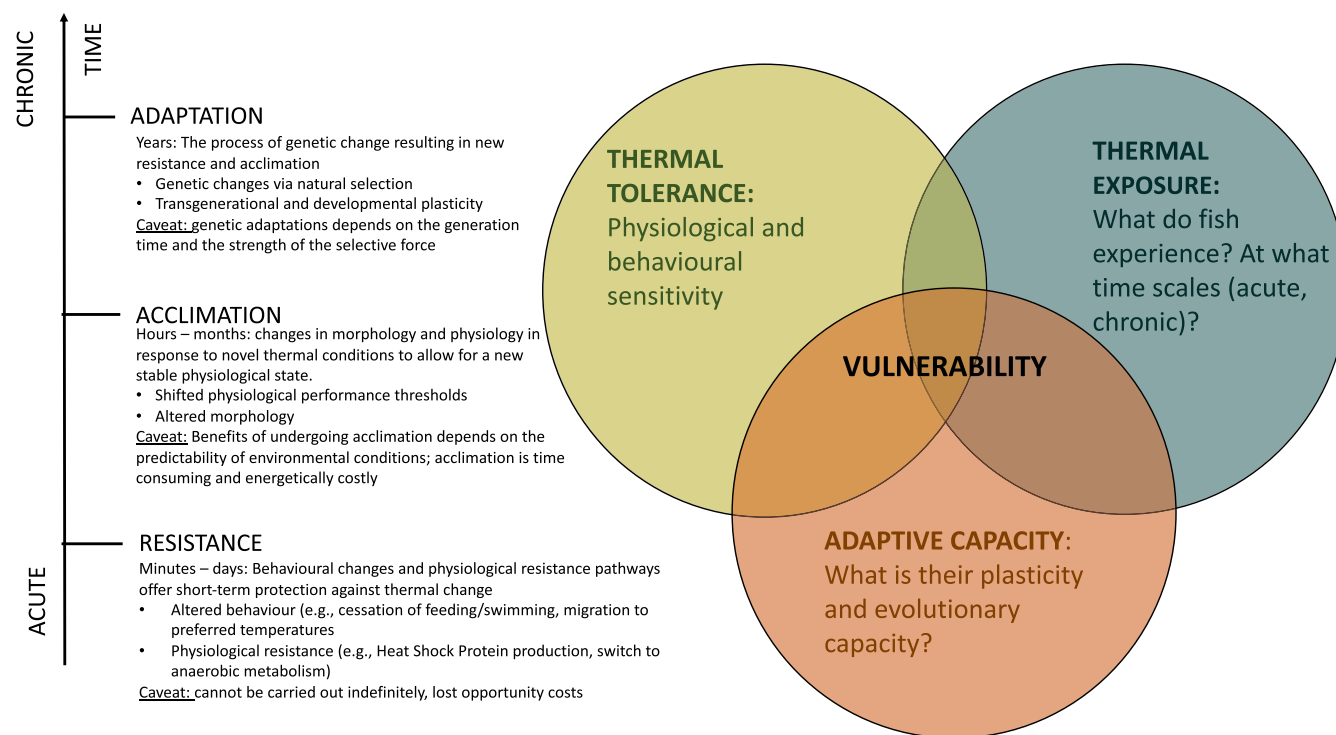
within a population of Pacific salmonids can occur at a rate that allows these fish to withstand climate-induced warming remains unknown (Angilletta et al., 2008). Traits (and the thermal metrics derived from them) related to thermal tolerance and performance have been found to be heritable in salmonids. Research into the genomic underpinnings for thermal tolerance in salmonids is underway; however, the genomic basis of thermal tolerance is highly complex and research into microevolution towards increased thermal tolerance is extremely limited (Chen et al., 2017). Muñoz et al. (2015) quantified the heritability of cardiac thermal performance. Chen et al. (2017)



**FIGURE 5** Critical thermal maximum (CT<sub>max</sub>) values obtained for coho salmon ( $n = 16$  for all test groups) when using different acclimation temperatures and heating rates (adapted from Becker & Genoway, 1979).

and demonstrated genomic patterns of evolutionary adaptation to extreme climates among populations of redband trout (*Oncorhynchus mykiss*) and identified potential candidate genes related to the thermal sensitivity of cardiac function (including IRX5, CASQ1, CACID and TITIN). However, thermal performance is a polygenic phenotype that will likely require whole genome sequencing to identify genomic regions that are associated with thermal limits (Chen et al., 2015). As research into population- and life stage-specific thermal tolerance develops, an understanding of how species upper thermal tolerance has evolved over time is a critical component in understanding how they may continue to adapt to future temperature conditions (Bennett et al., 2021). There are many behavioural and physiological mechanisms (i.e. behavioural thermo-regulation and aerobic scope) that may act synergistically to limit or promote the potential for adaptive evolution to warmer environmental conditions. Therefore, a consideration of the various UTL metrics that reflect these mechanisms will be central in making accurate predictions about the potential for adaptive evolution of thermal tolerance under future warming.

Successful evolutionary adaptation to new environmental conditions depends largely on how the rate of environmental change compares to the generation time of the species or population. Previous studies indicate that climate-induced freshwater warming acts as a relatively strong selective pressure on genes controlling migration timing and suggests that the rate of evolution towards earlier migration can be as short as two generations (Kovach et al., 2012; Reed et al., 2011). For traits related to thermal tolerance, there is evidence of intraspecific variation (Chen et al., 2013; Eliason et al., 2011; Eliason, Wilson,



**FIGURE 6** Framework for assessing vulnerability to warming in fishes. The three components of vulnerability are depicted in the Venn diagram. Represented along the vertical axis are definitions and temporal scales over which the components of thermal tolerance are most relevant.



et al., 2013; see Figure 3). However, in the research we reviewed, individual variation within a population appears to be limited (Chen et al., 2017; Muñoz et al., 2015). This lack of variation can be partially attributed to the small size of some populations of Pacific salmonids and their proclivity for natal homing which has allowed for evolution of thousands of populations that are finely tuned to local conditions. These findings suggest that a single population may not possess sufficient genetic variation to evolve their thermal tolerance rapidly enough to withstand expected future warming. In these cases, consideration of the heritable intraspecific variation in upper thermal tolerance is essential to determining whether adaptation could occur through the transfer of thermally tolerant traits from one population to another.

In wild populations, straying acts as a crucial evolutionary feature that can allow for adaptation over relatively short ecological time frames (Keefer & Caudill, 2014). Introduction of novel genetic diversity into a population can increase the probability of genetic rescue and allow for a population to more quickly adapt to new environmental conditions (Carlson et al., 2014). However, the transfer of traits that are maladaptive to local conditions is also a possibility, especially given the highly specific local adaptations that these populations have developed (Fraser et al., 2011). The potential for straying to impact the genetic structure of recipient populations depends on the relative size of the donor and recipient population as well as the frequency of straying from the donor population (Bett et al., 2016). Research into the extent and causes of straying suggests that straying becomes more common with increasing temperatures as fish are more likely to enter non-natal tributaries when main-stem temperatures reach stressful levels (Keefer, Caudill, et al., 2008; Westley et al., 2015). Therefore, as freshwater temperatures continue to rise in response to climate change, increased rates of straying between wild populations may occur. With respect to climate warming and thermal tolerance, strays have the potential to either hinder or enable evolution in thermal tolerance traits. For example, high rates of straying by the Chilko sockeye salmon population to the Gates Creek spawning area were observed in 2012 (Bett et al., 2016). Gates Creek is a relatively small population and interbreeding with Chilko sockeye salmon could allow for the transfer of exceptional cardiac adaptations that individuals from this donor population possess (Anttila et al., 2019; Eliason et al., 2011). Conversely, less favourable traits like lower fecundity, smaller eggs and low thermal tolerance during the embryonic life stage could also be transferred (Bett et al., 2016; Whitney et al., 2013). Therefore, building upon research into the extent and rate of straying in Pacific salmonids and relating this work to our current understanding of population and life stage-specific UTL will be an important component in determining the potential for Pacific salmon to successfully adapt to changing environmental conditions.

When endeavouring to assess vulnerability to warming temperatures, the range of responses (behavioural, physiological, biochemical, morphological, phenological) must be considered. Further, these responses have varying temporal importance (acute, chronic, inter-generational; Figure 6). This review has identified several approaches that aim to quantify thermal tolerance at the physiological ( $CT_{max}$ , Aerobic and Cardiac Scope, biochemical responses) and behavioural

(thermal migration barriers, preference/avoidance) levels. Each of these approaches provides meaningful information on the overall thermal response that fish will have as water temperatures rise (Bates & Morley, 2020). However, the results from these different approaches cannot be used interchangeably and applying these values to projections on future persistence and survival must be done with an understanding of what information each of these approaches offers. When selecting a UTL estimate to use in the development of thermal management guidelines, we must consider the UTL metrics in context; both the intended use (informing management tactics like fish quotas or fisheries closures, modelling future persistence) and limitations of what information each metric provides must be taken into account to avoid over- or underestimating vulnerability. For example,  $CT_{max}$  provides insights about UTLs on acute timescale and thus,  $CT_{max}$  values should not be used to investigate tolerance to chronic high temperature exposures, when other metrics are available. However, plasticity in the thermal response with acclimation or acclimatization that is seen in  $CT_{max}$  studies can provide useful information on the non-evolutionary adaptive capacity (plasticity) of a given population or species. Further, physiological resistance pathways, such as the production of heat shock proteins or a shift to anaerobic metabolism are important mechanisms for surviving acute thermal stress (Bates & Morley, 2020). With respect to more chronic temperature exposures, the ability to acclimate and extend thermal tolerance is more critical to survival. In these cases, using results from multiple approaches can provide meaningful insights regarding both the numeric temperature limits and the mechanisms governing them. For example, new molecular approaches can be integrated with other whole animal or organ-level physiological methods (i.e. aerobic or cardiac scope) to provide further insights into the mechanisms driving thermal mortality (Jeffries, Hinch, Martins, et al., 2012; von Biela et al., 2019). Given that UTL are shaped by a variety of physiological and behavioural responses there are instances where using a combination of metrics may be most appropriate and allow for a more robust representation of the overall thermal response to warming. Using multiple metrics in concert can also aid in determining whether a metric is either physiologically or ecologically relevant or both (two concepts that are often used interchangeably in the literature). When metrics assessing various behavioural or physiological aspects of the thermal response to warming yield consistent results, such metrics are extremely powerful in their possessing both ecological relevance (are indicative of how individuals may respond to warming temperatures) and physiological relevance (provide a physiological thermal limit beyond which a species can no longer survive and persist). Though sparse, consistent behavioural and physiological metrics are present in the literature in the form of thermal migration barriers and Pejus temperatures ( $T_p$ ). For example, thermal migration barriers have been identified for adult sockeye salmon from the Chilko, Quesnel and Stellako/Early Stuart populations at  $\sim 19^\circ\text{C}$  (MacDonald et al., 2010; Martins et al., 2011).  $T_p$  estimates for these same populations are within a similar range (Chilko:  $20.7^\circ\text{C}$ , Quesnel:  $18.5^\circ\text{C}$ , Stellako/Early Stuart:  $19.9^\circ\text{C}$ ; Eliason et al., 2011). Alternative efforts for reconciling UTLs derived from experiments with differing protocols have recently been proposed by

Ørsted et al. (2022) in which ecological and physiological evaluation of thermal stress is based off the net sum of two biological processes; thermal injury (disruption of homeostatic regulation) and thermal repair (temperatures where homeostatic regulation can be carried out).

## 4 | CONCLUSIONS

We outlined the importance of considering the potential sources of variation in reported UTL for Pacific salmon and demonstrated how the use of species- and life stage-specific UTLs can provide a more accurate depiction of how these species may respond to future warming. In instances where complimentary metrics are available (i.e. thermal performance curves for aerobic scope and cellular responses to thermal stress) they can and should be used in concert to develop a more robust understanding of UTL. Furthermore, metrics that are consistent across both physiological and behavioural studies are most powerful and possess high ecological relevance as they are rooted in a mechanistic understanding of why fish can no longer thrive at high temperatures and reflective of the thermal experience and energetic requirements of the fish in nature. Major data deficiencies exist at region/latitude, species, population and life stage levels for all species reviewed, however pink and chum salmon are vastly underrepresented in the literature.

We identified seven methodological approaches employed to quantify UTL, some of which possess more ecological relevance than others and, in certain cases, their use may be limited to specific life stages. Further, our findings demonstrated that both inherent (species, population, life stage) and methodological (within and among approaches) variation contribute to the considerable differences observed in proposed UTL—a finding that would be universally applicable to most fish species. Further, all of the approaches we reviewed possess specific limitations owing to logistics and/or ecological relevance and there is certainly room for the development of novel approaches that fill gaps in our understanding of the overall thermal response of Pacific salmonids to rising temperatures. Recent studies that have examined thermal stress through cellular stress responses (e.g. heat shock protein induction, transcriptomics), represent another promising methodological tool for identifying UTLs in Pacific salmon (Carey et al., 2019; Jeffries et al., 2014; Jeffries, Hinch, Sierocinski, et al., 2012; von Biela et al., 2020). Lastly, there is a dearth of information on the role of body size and sex in relation to thermal tolerance and is an area where future research should be directed. This finding is likely not just limited to Pacific salmonids and would be relevant to most fish species. Overall, future fisheries management and conservation actions should be developed with careful consideration of the nuances we identified in how thermal tolerance is assessed.

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## CONFLICT OF INTEREST STATEMENT

The authors have no conflict of interest to declare.

## DATA AVAILABILITY STATEMENT

All data are available in a supplemental file attached to this paper.

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## SUPPORTING INFORMATION

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